

Skewed contributions of individual trees to indirect nature experiences

Daniel T.C. Cox^{a,b,*}, Jonathan Bennie^c, Stefano Casalegno^a, Hannah L. Hudson^a, Karen Anderson^a, Kevin J. Gaston^a

^a Environment & Sustainability Institute, University of Exeter, Penryn, Cornwall TR10 9FE, UK

^b European Centre for Environment and Human Health, University of Exeter Medical School, Truro, Cornwall TR1 3HD, UK

^c Centre for Geography, Environment and Society, University of Exeter, Penryn, Cornwall TR10 9FE, UK

ABSTRACT

Exposure to nature is associated with a broad range of benefits to human health. Whilst there has been exploration of how these experiences vary amongst people, the converse – how different individual organisms contribute to human nature experiences – has largely been overlooked. The most common way that people experience nature occurs indirectly, when they are in a room with a natural view. Here, we estimate variation in how individual trees provide indirect nature experiences in an urban human population. As a proxy for its contribution towards indirect nature experiences, within an extended urban area in southern England, UK ($n = 612,920$) we calculated the number of buildings with line of sight to each tree. We then modelled each tree's contribution towards these experiences against potential predictors, namely tree height, land ownership, social deprivation, while controlling for human population density. We demonstrate that a small number of trees contribute disproportionately towards indirect nature experiences, with individual trees in socio-economically deprived high density housing falling within the viewscape of significantly more buildings. Further, trees in private gardens were generally more important for providing indirect nature experiences than those in public green spaces. This novel study demonstrates the skewed contribution of different organisms to human population indirect nature experiences. This approach can be applied more broadly to understand how individual organisms provide indirect, incidental and intentional nature experiences. Understanding the ecology behind human nature experiences is an important step towards linking urban design and policy for maximising the health benefits from nature.

1. Introduction

Urbanisation is emerging as one of the most important human health issues of the 21st century (World Health Organisation, 2015), with cities becoming epicentres for chronic and non-communicable physical and mental health conditions (Dye, 2008). Nature in cities has the potential to mitigate many of these health issues, with demonstrable links between exposure to nature and health and well-being benefits (e.g. Keniger, Gaston, Irvine, & Fuller, 2013). These benefits span a remarkable range of health outcomes, including but not limited to, reduced all-cause mortality and mortality from cardiovascular disease (e.g. Donovan et al., 2013; Mitchell & Popham, 2008), reduced healing times (Raanaas, Patil, & Hartig, 2012; Ulrich, 1984), reduced stress (e.g. Van Den Berg & Custers, 2011), reduced respiratory illness and allergies (e.g. Hanski, Herten, & Fyhrquist, 2012; Lovasi, Quinn, Neckerman, Perzanowski, & Rundle, 2008), improved self-reported wellbeing and reduced risk of poor mental health (e.g. Cox, Shanahan, Hudson, Plummer, et al., 2017; Dallimer et al., 2012; Fuller, Irvine, Devine-Wright, Warren, & Gaston, 2007), and improved cognitive ability (e.g. Berman, Jonides, & Kaplan, 2008).

Within the urban environment exposure to nature is more complex

and versatile than often portrayed; to a greater or lesser extent people are exposed to components of nature throughout their daily lives. Keniger et al. (2013) identified three main types of nature experience. First, people experience nature indirectly while not actually being present in it (e.g. having a view of nature from home or work). Second, nature is experienced incidentally while carrying out another activity (e.g. walking to the shops). Third, people intentionally experience nature (e.g. visiting parks or gardens). With the rise in urban living, most people now spend much of their day indoors, therefore the green viewscape from home or from work often constitutes by far their most common nature experience (Cox, Hudson, Shanahan, Fuller, & Gaston, 2017b). Having a room with a view of nature does not necessarily mean that people are continuously experiencing that view. Instead, people spend a significant amount of time with their attention directed towards specific tasks, and the presence of a window with a natural scene allows micro-restorative experiences (Kaplan, 1993, 2001), with scenes that are more fascinating being likely to be more restorative (Kaplan & Kaplan, 1989). Indeed, there is robust evidence to suggest that indirect nature experiences provide a broad range of health and wellbeing benefits, including increased psychological wellbeing (Kaplan, 2001), improved cognitive function (e.g. Zijlema et al., 2017) and

* Corresponding author at: Environment & Sustainability Institute, University of Exeter, Penryn, Cornwall TR10 9FE, UK.

E-mail addresses: d.t.c.cox@exeter.ac.uk (D.T.C. Cox), j.j.bennie@exeter.ac.uk (J. Bennie), stefano@casalegno.net (S. Casalegno), h.hudson@exeter.ac.uk (H.L. Hudson), karen.anderson@exeter.ac.uk (K. Anderson), k.j.gaston@exeter.ac.uk (K.J. Gaston).

<https://doi.org/10.1016/j.landurbplan.2019.01.008>

Received 18 June 2018; Received in revised form 21 November 2018; Accepted 11 January 2019

Available online 23 January 2019

0169-2046/ © 2019 The Authors. Published by Elsevier B.V. This is an open access article under the CC BY license (<http://creativecommons.org/licenses/by/4.0/>).

concentration (Bodin, Björk, Ardö, & Albin, 2015), reduced healing times (Ulrich, 1984) and reduced stress at work (Kaplan, 1993; Largo-Wight, Chen, Dodd, & Weiler, 2011; Leather, Pyrgas, Beale, & Lawrence, 1998). However, indirect nature experiences are not evenly distributed across the population, with a relatively small proportion of people spending disproportionately more time indirectly experiencing nature (Cox et al., 2017b).

As well as there being variation in the degree to which people experience nature, there will also be variation in how individual organisms contribute towards providing those experiences (Gaston et al., 2018). Some will almost certainly be major contributors, others minor ones and some may not contribute at all. However, to date this issue has not been explored. Intuitively, those individuals that are present where people occur and are more visible within natural viewsapes will be seen by a greater number of people, and so will be relatively more important for providing indirect nature experiences. For example, trees are a highly visible component of many urban viewsapes (Nowak et al., 1996) and potentially provide a wide range of health and well-being benefits (Salmond et al., 2016). As such, their location in the landscape seems likely to be critical for determining how often and for how long people experience them (Kardan et al., 2015). Those trees that fall within the viewscape of more buildings are likely to be disproportionately more important for providing health benefits associated with indirect nature experiences.

Here, we explore how individual urban trees vary in their contribution to indirect nature experiences in a human population. Determining which individual trees contribute more experiences may not be straightforward, but one might predict that larger trees are more visible, as are those that occur in areas of denser human population and on public lands. Trees in wealthier areas seem likely also to provide more nature experiences, because less dense and more designed urban spaces tend to favour views from associated properties (Landry & Chakraborty, 2009). We test these predictions using a spatial dataset derived from aerial photography and colour infrared data with high resolution digital surface models, within which the location of every tree within an extended urban area in southern England, UK was mapped, and its height estimated. Using spatial analysis algorithms implemented within geographic information systems (GIS), we determined the number of buildings that had line of sight to each tree, which we used as a proxy for indirect nature experiences from home or from work. To understand the important predictors of a tree's contribution towards these experiences, and so inform urban design and planning towards green health interventions, we modelled the response against potential predictors, namely: tree height, because this may influence a tree's visibility; landownership, because broadly speaking this will determine the type of land and tree management; and social deprivation to account for socio-economic variation in human neighbourhoods which is known to influence both tree and human populations (e.g. Ferguson, Roberts, McEachan, & Dallimer, 2018). We expect that there will be a positive relationship between the number of buildings with line of sight to a tree and human population density, therefore to tease out the effects of other predictors we controlled for the potentially confounding effects of spatially-variable human population density surrounding each tree.

2. Methods

2.1. Study area

This study focused on the urban area of the 'Cranfield triangle', a region in southern England, U.K. (52°07'N, 0°61'W). This comprises the three adjacent towns of Milton Keynes, Luton, and Bedford, which have a combined human population of c.546,000 (Office of National Statistics, 2016), and occupy c.197 km². Within this region there is great variation in human population density and urban form (including representatives of low- and high density living). The urban limits were

defined as where continuous residential or commercial properties ended, and rural green space began (Gaston, Warren, Thompson, & Smith, 2005).

2.2. Data sources

A fine-spatial resolution digital surface model (DSM) for the study area was generated from airborne light detection and ranging (LiDAR) data captured by the Natural Environment Research Council Airborne Research and Survey Facility between June and September 2012. The DSM has a horizontal resolution of 1 m and widely known biases for vegetation height retrieval that result in a nominal vertical accuracy of ± 0.5 m. The position (centre point) and height of every tree > 3 m in height within a 100 m buffer surrounding the urban limits of the Cranfield triangle was obtained as a point vector point data from a commercial product called the "National Tree Map" (NTM; $n = 612,920$; Bluesky International LTD; <http://www.blueskymapshop.com>). A vector polygon layer of the outlines of buildings and structures was obtained from the United Kingdom Ordnance Survey VectorMap (OS VectorMap; <http://digimap.edina.ac.uk/>; accessed 2016). All building polygons with an area between 36 m² and 360 m² were selected; this was a size range that was considered to represent buildings with windows that people are likely to use for home and work (i.e. removing smaller non-inhabited structures such as sheds and garages, and larger structures such as warehouses). All spatial data manipulations were performed in R software for statistical computing version 3.3.3 (R Development Core Team, 2016) and QGIS v2.14 (Quantum GIS Development Team, 2016). The registration correspondence between the NTM and the LiDAR DSM was visually good, but we were unable to arrive at a spatially-distributed registration error estimate for two reasons. First the NTM was a commercial product, for which quantitative error information were unavailable beyond a generic statement from Bluesky that states "overall canopy coverage represented in NTM is accurate to over 90% and over 95% accurate within 50 m of buildings". Second, the LiDAR data had an associated measurement uncertainty that was spatially variable but unknown for this site – the technical specifications of the Leica ALS500-II sensor state that it has a "lateral placement accuracy of 7–64 cm and vertical placement accuracy of 8–24 cm (one standard deviation) from full-field-filling targets" (www.nts-info.com/inventory/images/ALS50-II.Ref.703.pdf).

2.3. Identifying lines of sight between buildings and trees

As a measure of the contribution of individual trees to indirect nature experiences, we calculated the number of buildings that had the potential for an unobscured view of each tree from an upstairs window, allowing for the effects of topography, other buildings and structures and vegetation in obscuring this line of sight. A tree will be perceived to be smaller if it is further away, and as a consequence the ability of a viewer to distinguish ecological detail will lessen, which is likely to influence the type of nature experience. For example, it is possible to recognise a small songbird such as a robin *Erithacus rubecula* as a perching bird, by eye at 350 feet (c.100 m) but not to determine the species (Wood, 1937). At a distance of 100 m the perceived height of an averagely sized urban tree (8.36 m, min 1.0 m; max 49.7 m) will be 0.0836 m (true height/distance = perceived height/distance to perceived height, or in this case; $(8.36/100) \times 1$). We considered that at sizes much smaller than this features of the tree will be difficult to distinguish and so the type of nature experience is likely to be different. Therefore, to capture the likely near nature experience we considered a tree to be potentially "visible" from a building if (a) it was within 100 m of at least one edge of the building (although we recognise that trees further than 100 m away may also be visible from buildings), and (b) there was a clear line of sight from the central point of the tree (at mid-height) to the edge of the building (at one metre below the height of the building at that edge, to allow for the view from a top floor window

slightly below the roof-line). The DSM, which includes terrain, vegetation canopy, buildings, and structures such as walls, was used to determine where lines of sight were obscured. We calculated the number of buildings that had line of sight to each tree, as a proxy for a tree's contribution towards indirect nature experiences. This was done using an algorithm written in R (R Development Core Team, 2016) using the raster package (Hijmans, 2016) and rgdal package (Bivand, Keitt, & Rowlingson, 2016). The algorithm first loops through each tree in the vector file, and selects all buildings within a 100 m radius. Each building is then selected in turn, and all DSM pixels falling on the edge of the building (i.e. intersecting the building outline) are selected. The DSM is then checked for a direct line of sight between the central point of the tree and the building edge, using the method described in Bennie, Davies, Inger, and Gaston (2014). To prevent the line of sight to the centre of a tree being apparently blocked by the outer branches of the tree itself, or by mismatches between the polygon edge of the building and the height of the DSM, obstructions within three metres of the centre of the tree or edge of the building were ignored. For each building within 100 m that had an unobstructed line of sight to a tree, a tally was added to a column in the attribute table of the tree point data. In this way, the number of buildings from which each tree was “visible” could be counted.

2.4. Land ownership, social deprivation and human population density

We determined land ownership for each tree by again using the OS VectorMap. We created a three-level factor, of non-neighbourhood trees, neighbourhood trees on private land, and neighbourhood trees on public land (termed land ownership). Non-neighbourhood trees were those without line of sight to a building ($n = 231,783$). We then developed a spatial layer to categorise remaining trees as being located within a residential garden polygon ($n = 177,842$), or being located on publicly owned land ($n = 202,294$).

To arrive at a generalised measure of deprivation in the neighbourhood surrounding each tree, we used weekly household wages. These were derived from model-based estimates for households (Office of National Statistics; <http://www.neighbourhood.statistics.gov.uk>). This index estimates income per household per week in pounds sterling, from data identified during the period April 2007 to April 2008. The household data are averaged across the Middle layer Super Output Area (MSOA), a geographical hierarchy consisting of 2000–6000 households and were the most recent data currently available. To each tree, we assigned the weekly household income from the MSAO that the tree was located within.

Based on the UK gridded population map (Reis et al., 2016), we estimated human population density as the number of people within a 500 m radius surrounding each tree. Estimating human population density at this scale provides an estimate of high to low density housing surrounding a tree, without being skewed by localised clustering of buildings. This dataset consists of gridded population data with a spatial resolution of 1 km^2 , assigned to the UK National Grid. We scaled the human population density for each tree, and where the 500 m radius covered multiple grid cells we weighted this population density by the percentage of the radius in each cell. The human population density in the vicinity of each tree was only weakly correlated to the response (i.e. the number of buildings that had line of sight to each tree; Pearson's correlation coefficient = 0.43).

2.5. Statistical analysis

We log-transformed the response variable (i.e. the number of buildings that had line of sight to each tree), so that it was approximately normally distributed. We used the *dnearneigh* function in the *spdep* package (Bivand & Piras, 2015) to create spatial weights for neighbours list of the trees that fell within 100 m of each tree (mean = 114; range = 1–578). Trees with no neighbouring trees within

100 m were excluded ($n = 5$). We then built a spatially lagged dependent variable model (also known as an SLX model) using the *lmSLX* command in the ‘spdep’ R package. Spatially lagged models such as SLX provide coefficients for the effect of independent variables on the response (‘direct’ effect), and coefficients for the mean effect of the values of the independent variables at the sites of neighbouring trees (‘indirect’ effects). For example, in this case, along with modelling the direct influence of tree height on the response, the SLX model also accounts for an indirect effect of mean height of neighbouring trees. We used the *impacts* command in the ‘spdep’ package to give the overall effect of both the direct and indirect effects (i.e. total coefficients). The model took the following form:

$$\gamma = \chi\beta + \omega\chi\theta + \varepsilon$$

where γ is the number of buildings with line of sight to that tree. $\chi\beta$ represents the independent variables and coefficients, namely: tree height (predictor), neighbourhood income (predictor), land ownership (predictor) and human population density (confounding). The spatial weights matrix is $\omega\chi\theta$, which is the average value of independent variables of trees within 100 m of the response and ε is the spatially independent error.

3. Results

Within the urbanised study area (196.7 km^2) there were 612,920 trees (Milton Keynes, 308,501 in 104.8 km^2 ; Luton, 196,365 in 58.3 km^2 ; Bedford, 108,054 in 33.6 km^2). This is equivalent to 1.12 trees per person, with an average human population density of 87 (± 63) people within 100 m of each tree. The contribution of each tree to indirect nature experiences (i.e. the number of buildings that had line of sight to that tree) varied across the three towns (Fig. 1), and was highly skewed, with 75% of all indirect experiences being provided by c.25% of trees (Fig. 2a). Trees had an average height of $8.5 \pm 4.8 \text{ m}$, with non-neighbourhood trees being $10.2 \pm 5.5 \text{ m}$ tall (i.e. those with line of sight to no buildings), and neighbourhood trees on public land ($8.4 \pm 4.3 \text{ m}$) tending to be taller than those on private land ($6.5 \pm 3.4 \text{ m}$). 380,137 (62%) trees fell within the viewscape of at least one building; of these 202,294 (53%) were located on public land and could be seen by 11 ± 13 buildings, while 177,843 (47%) trees were located in private gardens and could be seen by 24 ± 17 buildings.

The direct and indirect predictors in the SLX model explained a high proportion of the variance, with a pr^2 of 0.81 (Table 1). When direct and indirect effects are considered together, the height of a tree was significantly, but weakly negatively associated with the number of buildings that could see it ($-0.02 (\pm 3.8\text{e}-4)^{***}$; Table 1). When considering direct effects alone, there was a significant, weak positive relationship between the response and the height of a tree ($0.013 (\pm 2.4\text{e}-4)^{***}$; Table 1), whilst for indirect effects, there was a significant weak negative relationship with the mean height of neighbouring trees ($-0.033 (\pm 0.001)^{***}$; Table 1).

Direct and indirect effects considered both separately and together showed that, of neighbourhood trees those located in private gardens were more important for indirect nature experiences than those located on public land (Table 1; Fig. 2b). Further, trees in more socio-economically deprived areas contributed more to indirect nature experiences than those in wealthier neighbourhoods (Table 1). As expected, when direct and indirect effects were considered together, there was a positive relationship between the contribution of individual trees to indirect nature experiences and human population density ($65.6 (\pm 0.6)^{***}$; Table 1; Fig. 2c). However, on the basis of direct effects alone, there was a weak negative relationship between the response and human population density ($-82.0 (\pm 7.3)^{***}$; Table 1), but a strong positive relationship with the mean population density of the tree's neighbours ($147.6 (\pm 7.4)^{***}$; Table 1).

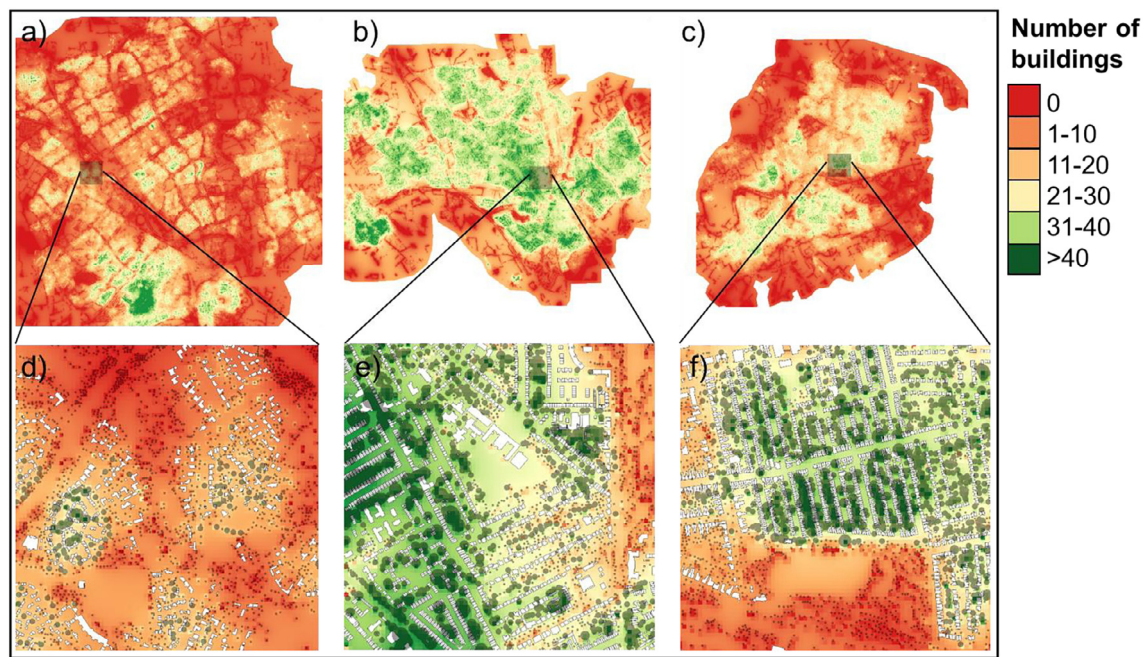


Fig. 1. Spatial variation in the contribution of individual trees to indirect nature experiences in three towns in Southern England: (a) Milton Keynes, (b) Luton, and (c) Bedford. Heat maps have been graded from red (areas containing trees with line of sight to no buildings) to dark green (areas containing trees with line of sight to > 40 buildings). (d–f) magnified area within each town, illustrating typical urban forms. Green circles show trees, with the size of each circle being weighted by the number of buildings with line of sight to that tree. Upper panels have a width of 12,000 m, and lower panels 750 m. (For interpretation of the references to colour in this figure legend, the reader is referred to the web version of this article.)

4. Discussion

This novel study demonstrates for the first time the importance of considering the spatial and volumetric role of specific nature components for providing indirect nature experiences to people. We show that the contribution of individual trees to indirect nature experiences is highly skewed. Quantifying the relative importance of individual trees to people, and understanding how trees are indirectly experienced is an important step towards linking urban design and policy for maximising the health benefits from urban nature.

Trees located in private gardens are generally visible from more buildings than those neighbourhood trees on public land. When considering direct and indirect neighbourhood effects together, there was a weak negative relationship between a tree's height and its importance for providing indirect nature experiences. In retrospect this is unsurprising considering that trees that were visible from at least one building were on average 2.9 m shorter than trees with no line of sight to a building. Further, direct effects show that if a tree is tall, while its neighbours are short, it can also be seen by more buildings. This suggests that although in the main, it is a tree's location that is the most

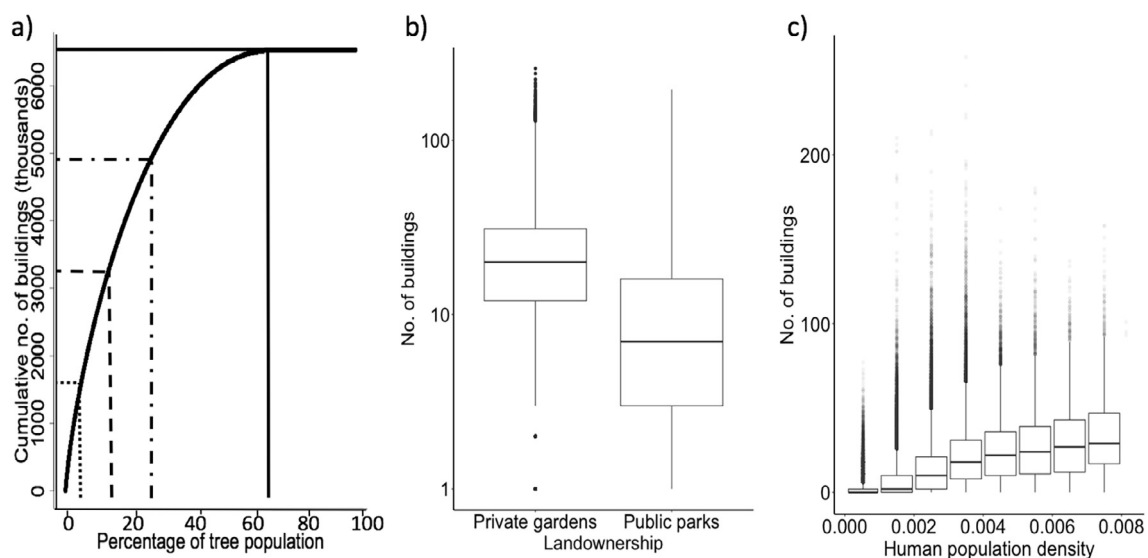


Fig. 2. (a) The cumulative number of buildings that have line of sight to each tree ($n = 612,920$). We show the percentage of the tree population that accounts for 25% (dotted line), 50% (dashed line), 75% (dash/dot line) and 100% (solid line) of total indirect experiences (individual buildings may have line of sight to multiple trees, therefore the number of buildings shown here is an over representation of the true number at the study site). Relationships between the contribution of individual trees to indirect nature experiences, and (b) land ownership (*Note: we do not show non-neighbourhood trees because these were selected as having line of sight to no buildings*), and (c) human population density.

Table 1

Predictors of the contribution of individual urban trees to indirect nature experiences in an urban human population. We show the direct (effect of independent variables on the response), indirect (mean effect of independent variables on neighbouring trees) and total (direct and indirect combined) unstandardized coefficients. Land ownership is against a base factor level of non-neighbourhood trees, and statistical significance is shown (* $p < 0.05$; ** $p < 0.01$; *** $p < 0.001$; see supplementary material for the standardised coefficients (Table S1) and bivariate coefficients (Table S2)). The pseudo r^2 squared is McFadden's.

Variable	Direct coefficients	Indirect coefficients	Total coefficients
Tree height (m)	0.013 ($\pm 2.4e-4$)***	-0.033 (± 0.001)***	-0.02 ($\pm 3.8e-4$)***
Human population density [#]	-82.0 (± 7.3)***	147.6 (± 7.4)***	65.6 (± 0.6)***
Income ^{##}	-1.4e-4 ($\pm 2.3e-5$)***	-5.8e-4 ($\pm 3.0e-5$)***	-7.2e-4 ($\pm 6.0e-6$)***
<i>Land ownership</i>			
Private garden	2.0 ($\pm 3.8e-3$)***	1.3 ($\pm 5.5e-3$)***	3.3 ($\pm 3.8e-3$)***
Public land	1.7 ($\pm 3.2e-3$)***	0.2 ($\pm 5.1e-3$)***	1.9 ($\pm 3.9e-3$)***
pR ²	0.81		

[#] Number of people per 500 m radius of each tree.

^{##} Average weekly household income.

important determinant of their provision of indirect nature experiences, individual characteristics of the tree such as height can further influence this provision. The management of garden trees by individual households has the potential to influence the nature experiences of many more households in the neighbourhood. Unfortunately, largely due to conflict with urban intensification these trees may also be at the greatest risk of removal (Wyse, Beggs, Burns, & Stanley, 2015). Therefore conservation schemes need to raise awareness of the importance of garden trees not only for the health and well-being of the household whose land the tree is on, but also of their neighbours who benefit from it through indirect nature experiences. As expected when considering direct and indirect neighbourhood effects together, there was a positive relationship between human population density and the provision of nature experiences, however individually trees with more space around them (i.e. lower human population density) had line of sight to more buildings (see Human population density, Table 1). Those trees located in low income, high density housing tended to fall within the viewscape of significantly more buildings, and thus have the potential to provide indirect nature experiences to more people. These trees are often located in areas with low levels of green space where people already have a reduced daily exposure to nature (Shanahan, Lin, Gaston, Bush, & Fuller, 2014), therefore the importance of these trees may increase further as they contribute a greater proportion of a person's daily nature experience and so associated health and well-being benefits.

Undoubtedly urban form is critical for determining how often and for how long individual organisms are experienced by people in towns and cities. Although Milton Keynes, a planned green town, had more trees, individually those trees contributed less to indirect nature experiences than did trees in Luton and Bedford (Fig. 1). At face value this suggests that individual trees in Milton Keynes are less important, however, experiences of nature are of course more complex and diverse than providing only indirect nature experiences between individual trees and people. People who experience multiple trees simultaneously may have an enhanced experience, further trees also provide nature experiences incidentally, such as while people are travelling to work or the shops, or intentionally such as when people experience trees by going to public parks (Keniger et al., 2013). Trees also provide vital ecosystem services such as pollution filtration, storm water processing and thermoregulation, the effectiveness of which will often also be dependent on their abundance and the structure of the landscape (e.g. Endreny et al., 2017). Here we explored the contribution of individual trees to indirect experiences in English towns that contain a broad range of urban forms, with examples ranging from highly industrialised to planned green town suburbs. It is likely that the general patterns in the distribution of nature experience provision shown here will be applicable more broadly across different urban designs. For example, suburban trees in Tokyo (typified by highly compact urban design) are likely to be experienced by a greater number of people than suburban

trees in Brisbane (typified by sprawling green urban design).

Understanding how trees are experienced, both individually and collectively, is critical so that urban areas can be designed to maximise both positive nature experiences and ecosystem service provision, to mitigate many of the health issues associated with urban living while also maintaining the health benefits from living in these areas. Encouragingly, only 2.3% of buildings in the study area had no line of sight to a tree (within 100 m of the building). These buildings were generally sporadically distributed, although there was an example of a development with few indirect nature experiences of this kind (Fig. 3a). We show how the strategic positioning of a small number of trees in existing green spaces has the potential to provide indirect nature experiences to a disproportionate number of buildings (Fig. 3b). We recognise that some urban trees may deliver more nuanced ecosystem services – e.g. some species will blossom more prolifically than others, some will provide enhanced branching structures within which birds can nest, and others will deliver fruit or food sources for both humans and wildlife. Accounting for these spatially variable attributes and placing different values on trees accordingly could provide a further interesting avenue of future research. Thus, we posit that trees are a significant component of green infrastructure, and their position in the landscape has the potential not only to provide indirect nature experiences, but also incidental experiences and to promote people seeking intentional nature experiences (Beery et al., 2017; Church, 2018). Indeed, there is a positive association between neighbourhood tree cover and a person's orientation towards nature (Shanahan et al., 2017), with evidence that a person's nature orientation is linked not only to their desire to seek health benefits but also their ability to receive these benefits (Capaldi, Dopko, & Zelenski, 2014).

In sum, the biological world is hugely complex, and logically there will be significant variation not only in how people indirectly experience individual organisms, but also in how they are experienced incidentally and intentionally (Gaston et al., 2018). An organism's contribution to these experiences of nature will be dependent on its position in space and time relative to people, and in urban areas this will often be driven by urban form. For stationary organisms, urban forms with increased connectivity for people will increase the number of people that encounter them, while greater green space connectivity will increase the ability of mobile organisms such as birds to move within and between green spaces and so encounter people (Cox, Inger, Hancock, Anderson, & Gaston, 2016). Finally, a species' ecology will influence how conspicuous it is to human senses, with those species that are more visible or vocal, larger or have a stronger scent being more likely to be experienced by people. Future studies need to move beyond considering urban environments as binary combinations of green space and non-green space, and towards understanding the roles that individual organisms play in providing nature experiences. As urban intensification continues, while increased knowledge of the importance of greenspace is encouraging retrofitting of green infrastructure, relevant



Fig. 3. (a) An example of an urban form where few buildings have line of sight to trees. Black polygons show buildings with line of sight to no trees, white polygons show buildings with line of sight to at least one tree. Green circles show trees, with the size of the circle being weighted by the number of buildings with line of sight to that tree (0 (smallest); 1–2; 2–11; > 11 (largest)). Light grey polygons show existing public green spaces. (b) Example of how the strategic placement of three trees (red circles) in existing public green spaces can provide indirect nature experiences to occupants of 126 buildings (red polygons). (For interpretation of the references to colour in this figure legend, the reader is referred to the web version of this article.)

stakeholders need to invest often limited resources and space towards those species and groups of species that are experienced by the greatest numbers of people.

Acknowledgements

The research leading to this paper was supported by the “Fragments, functions and flows in urban ecosystems” project (NE/J015237/1) funded under the NERC Biodiversity and Ecosystem Services Sustainability (BESS) thematic program.

Conflicts of interests

We have no competing interests to declare.

Ethical clearance

Not applicable.

Appendix A. Supplementary data

Supplementary data to this article can be found online at <https://doi.org/10.1016/j.landurbplan.2019.01.008>.

References

- Beery, T. H., Raymond, C. M., Kyttä, M., Olafsson, A. S., Plieninger, T., Sandberg, M., et al. (2017). Fostering incidental experiences of nature through green infrastructure planning. *Ambio*, 46, 717–730. <https://doi.org/10.1007/s13280-017-0920-z>.
- Bennie, J., Davies, T. W., Inger, R., & Gaston, K. J. (2014). Mapping artificial lightscapes for ecological studies. *Methods in Ecology and Evolution*, 5, 534–540. <https://doi.org/10.1111/2041-210X.12182>.
- Berman, M. G., Jonides, J., & Kaplan, S. (2008). The cognitive benefits of interacting with nature. *Psychological Science*, 19, 1207–1212. <https://doi.org/10.1111/j.1467-9280.2008.02225.x>.
- Bivand, R., Keitt, T., & Rowlingson, B. (2016). *rgdal: Bindings for the 'Geospatial' data abstraction library. R package version 1.1-10*. <https://CRAN.R-project.org/package=rgdal>.
- Bivand, R., & Piras, G. (2015). Comparing implementations of estimation methods for spatial econometrics. *Journal of Statistical Software*, 63, 1–36. <http://www.jstatsoft.org/v63/i18/>.
- Bodin, T., Björk, J., Ardö, J., & Albin, M. (2015). Annoyance, sleep and concentration problems due to combined traffic noise and the benefit of quiet side. *International Journal of Environmental Research and Public Health*, 12, 1612–1628. <https://doi.org/10.3390/ijerph120201612>.
- Capaldi, C. A., Dopko, R. L., & Zelenski, J. M. (2014). The relationship between nature connectedness and happiness: A meta-analysis. *Frontiers in Psychology*, 5, 976. <https://doi.org/10.3389/fpsyg.2014.00976>.
- Church, S. P. (2018). From street trees to natural areas: Retrofitting cities for human connectedness to nature. *Journal of Environmental Planning and Management*, 61, 878–903. <https://doi.org/10.1080/09640568.2018.1428182>.
- Cox, D. T. C., Hudson, H. L., Shanahan, D. F., Fuller, R. A., & Gaston, K. J. (2017b). The rarity of direct experiences of nature in an urban population. *Landscape and Urban Planning*, 160, 79–84. <https://doi.org/10.1016/j.landurbplan.2016.12.006>.
- Cox, D. T. C., Inger, R., Hancock, S., Anderson, K., & Gaston, K. J. (2016). Movement of feeder-using songbirds: The influence of urban features. *Scientific Reports*, 6, 37669. <https://doi.org/10.1038/srep37669>.
- Cox, D. T. C., Shanahan, D. F., Hudson, H. L., Plummer, K. E., Siriwardena, G. M., Fuller, R. A., et al. (2017a). Doses of neighbourhood nature: Benefits for mental health of living with nature. *BioScience*, 67, 147–152. <https://doi.org/10.1093/biosci/biw173>.
- Dallimer, M., Irvine, K. N., Skinner, A. M. J., Davies, Z. G., Rouquette, J. R., Maltby, L. L., et al. (2012). Biodiversity and the feel-good factor: Understanding associations between self-reported human well-being and species richness. *BioScience*, 62, 47–55. <https://doi.org/10.1525/bio.2012.62.1.9>.
- Donovan, G. H., Butry, D. T., Michael, Y. L., Prestemon, J. P., Liebhold, A. M., Gatzliolis, D., et al. (2013). The relationship between trees and human health evidence from the spread of the emerald ash borer. *American Journal of Preventive Medicine*, 44, 139–145. <https://doi.org/10.1016/j.amepre.2012.09.066>.
- Dye, C. (2008). Health and urban living. *Science*, 319, 766–769. <https://doi.org/10.1126/science.1150198>.
- Endreny, T., Santagata, R., Perna, A., De Stefano, C., Rallo, R. F., & Ulgiati, S. (2017). Implementing and managing urban forests: A much needed conservation strategy to increase ecosystem services and human well-being. *Ecological Modelling*, 360, 328–335. <https://doi.org/10.1016/j.ecolmodel.2017.07.016>.
- Ferguson, M., Roberts, H. E., McEachan, R. R. C., & Dallimer, M. (2018). Contrasting distributions of urban green infrastructure across social and ethno-racial groups. *Landscape and Urban Planning*, 175, 136–148. <https://doi.org/10.1016/j.landurbplan.2018.03.020>.
- Fuller, R. A., Irvine, K. N., Devine-Wright, P., Warren, P. H., & Gaston, K. J. (2007). Psychological benefits of greenspace increase with biodiversity. *Biology Letters*, 3, 390–394. <https://doi.org/10.1098/rsbl.2007.0149>.
- Gaston, K. J., Soga, M., Duffy, J. P., Garrett, J. K., Gaston, S., & Cox, D. T. C. (2018). Personalised ecology. *Trends in Ecology and Evolution*, 33, 916–925. <https://doi.org/10.1016/j.tree.2018.09.012>.
- Gaston, K. J., Warren, P. H., Thompson, K., & Smith, R. M. (2005). Urban domestic gardens (IV): The extent of the resource and its associated features. *Biodiversity and Conservation*, 14, 3327–3349. <https://doi.org/10.1007/s10531-004-0513-6>.
- Hanski, I., Herten, L., & Fyhrquist, N. (2012). Environmental biodiversity, human microbiota, and allergy are interrelated. *Proceedings of the National Academy of Sciences of the United States of America*, 109, 8334–8339. <https://doi.org/10.1073/pnas.1205624109>.
- Hijmans, R. J. (2016). *Raster: Geographic data analysis and modelling. R package version 2.5-8*. <https://cran.r-project.org/package=raster>.
- Kaplan, R. (1993). The role of nature in the context of the workplace. *Landscape and Urban Planning*, 26, 193–201. [https://doi.org/10.1016/0169-2046\(93\)90016-7](https://doi.org/10.1016/0169-2046(93)90016-7).
- Kaplan, R. (2001). The nature of the view from home – Psychological benefits. *Environmental Behaviour*, 33, 507–542. <https://doi.org/10.1177/00139160121973115>.
- Kaplan, R., & Kaplan, S. (1989). *The experience of nature: A psychological perspective*. Cambridge: Cambridge University Press.
- Kardan, O., Gozhyra, P., Misic, B., Moola, F., Palmer, L. J., Paus, T., et al. (2015). Neighbourhood greenspace and health in a large urban centre. *Scientific Reports*, 5, 11610. <https://doi.org/10.1038/srep11610>.
- Keniger, L. E., Gaston, K. J., Irvine, K. N., & Fuller, R. A. (2013). What are the benefits of interacting with nature? *International Journal of Environmental Research and Public Health*, 10, 913–935. <https://doi.org/10.3390/ijerph10030913>.
- Landry, S. M., & Chakraborty, J. (2009). Street trees and equity: Evaluating the spatial distribution of an urban amenity. *Environment and Planning A*, 41, 2651–2670. <https://doi.org/10.1068/a41236>.

- Largo-Wight, E., Chen, W. W., Dodd, V., & Weiler, R. (2011). Healthy workplaces: The effects of nature contact at work on employee stress and health. *Public Health Reports*, 126, 124–130. <https://doi.org/10.1177/00333549111260S116>.
- Leather, P., Pyrgas, M., Beale, D., & Lawrence, C. (1998). Windows in the workplace – Sunlight, view, and occupational stress. *Environment and Behaviour*, 30, 739–762. <https://doi.org/10.1177/001391659803000601>.
- Lovasi, G. S., Quinn, J. W., Neckerman, K. M., Perzanowski, M. S., & Rundle, A. (2008). Children living in areas with more street trees have lower prevalence of asthma. *Journal of Epidemiology and Community Health*, 62, 647–649. <https://doi.org/10.1136/jech.2007.071894>.
- Mitchell, R., & Popham, F. (2008). Effect of exposure to natural environment on health inequalities: An observational population study. *Lancet*, 372, 1655–1660. [https://doi.org/10.1016/S0140-6736\(08\)61689-X](https://doi.org/10.1016/S0140-6736(08)61689-X).
- Nowak, D. J., Rowntree, R. A., McPherson, E. G., Sisinni, S. M., Kerkmann, E. R., & Stevens, J. C. (1996). Measuring and analyzing urban tree cover. *Landscape and Urban Planning*, 36, 49–57. [https://doi.org/10.1016/S0169-2046\(96\)00324-6](https://doi.org/10.1016/S0169-2046(96)00324-6).
- Office of National Statistics. (2016). 2011 Census Aggregate Data. UK Data Service. Edition June 2016. <http://doi.org/10.5257/census/aggregate-2011-1>.
- Quantum GIS Development Team. (2016). Quantum GIS Geographical Information System v.2.6 Open Source Geospatial Foundation project. Available online: <http://qgis.osgeo.org> [Accessed on 8 May 2016].
- R Development Core team. (2016). R: A language and environment for statistical computing. Vienna, Austria: R Foundation for Statistical Computing. Available online: <http://R-project.org> [Accessed on 8 May 2016].
- Raanaas, R. K., Patil, G. G., & Hartig, T. (2012). Health benefits of a view of nature through the window: A quasi-experimental study of patients in a residential rehabilitation center. *Clinical Rehabilitation*, 26, 21–32. <https://doi.org/10.1177/0269215511412800>.
- Reis, S., Steinle, S., Carnell, E., Leaver, D., Vieno, M., Beck, R., et al. (2016). *UK gridded population based on census 2011 and land cover map 2007*. <https://catalogue.ceh.ac.uk/documents/61f10c74-8c2c-4637-a274-5fa9b2e5ce44>.
- Salmund, J. A., Tadaki, M., Vardoulakis, S., Arbuthnott, K., Coutts, A., Demuzere, M., et al. (2016). Health and climate related ecosystem services provided by street trees in the urban environment. *Environmental Health*, 15(Suppl. 1), S36. <https://doi.org/10.1186/s12940-016-0103-6>.
- Shanahan, D. F., Cox, D. T. C., Fuller, R. A., Hancock, S., Lin, B. B., Anderson, K., et al. (2017). Variation in experiences of nature across a gradient of tree cover in compact and sprawling cities. *Landscape and Urban Planning*, 157, 231–238. <https://doi.org/10.1016/j.landurbplan.2016.07.004>.
- Shanahan, D. F., Lin, B. B., Gaston, K. J., Bush, R., & Fuller, R. A. (2014). Socio-economic inequalities in access to nature on public and private lands: A case study from Brisbane, Australia. *Landscape and Urban Planning*, 130, 14–23. <https://doi.org/10.1016/j.landurbplan.2014.06.005>.
- Ulrich, R. S. (1984). View through a window may influence recovery from surgery. *Science*, 224, 420–421.
- Van Den Berg, A. E., & Custers, M. H. G. (2011). Gardening promotes neuroendocrine and affective restoration from stress. *Journal of Health Psychology*, 16, 3–11. <https://doi.org/10.1177/1359105310365577>.
- Wood, H. B. (1937). How far can a bird be seen? *Auk*, 54, 96–97.
- World Health Organisation (2015). *Urban health*. [Accessed on 11th November 2016] http://who.int/topics/urban_health/en/.
- Wyse, S. V., Beggs, J. R., Burns, B. R., & Stanley, M. C. (2015). Protecting trees at an individual level provides insufficient safeguard for urban forests. *Landscape and Urban Planning*, 141, 112–122. <https://doi.org/10.1016/j.landurbplan.2015.05.006>.
- Zijlema, W. L., Triguero-Mas, M., Smith, G., Cirach, M., Martinez, D., Davdand, P. M., et al. (2017). The relationship between natural outdoor environments and cognitive functioning and its mediators. *Environmental Research*, 155, 268–275. <https://doi.org/10.1016/j.envres.2017.02.017>.